# POST-DELISTING MONITORING of the TINIAN MONARCH (Monarcha takatsukasae) <br> FINAL REPORT - June 2021 <br> Fred A. Amidon ${ }^{1}$, Ann P. Marshall ${ }^{1}$, Paul Radley ${ }^{2,4}$, and Eric VanderWerf ${ }^{3}$ 



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#### Abstract

The Tinian Monarch (Monarcha takatsukasae), locally known as Chichirikan Tinian, is a small forest bird endemic to the island of Tinian in the Commonwealth of the Northern Mariana Islands (CNMI). The species was listed as endangered in 1970 under the Endangered Species Act because its numbers were believed to be critically low as a result of forest loss. Survey data from 1982 indicated that the monarch population had increased or was at least stable and that the primary listing factor, loss of forest habitat, had been ameliorated. The Tinian Monarch was consequently downlisted in 1987 and delisted in 2004. Monitoring conducted between 2006 and 2011, as part of a 2005 USFWS Post-delisting Monitoring Plan for the species included roadside surveys, small-scale early warning plots, and land-use monitoring. Island-wide population surveys were additionally conducted in 2008 and 2013. Results of the roadside surveys indicate a significant decline in counts of Tinian Monarchs from 1999 to 2010, which was consistent with reported island-wide declines from 1996 to 2008 . However, increases in island-wide counts in 2013 indicate that the population may be fluctuating temporally. Data from the small-scale plot monitoring indicated an annual adult survival rate of $78 \%$, higher territory turnover rates for females ( $42 \%$ ) than males ( $35 \%$ ), and an overall territory occupancy rate of $94 \%$. Land-use monitoring, based on satellite imagery and CNMI land clearing permits, indicated an average of 20 hectares of land cleared per year from 2007 to 2010. An unexpected outcome of the monitoring was the observation and identification of avian pox in Tinian Monarchs but its impact on the species is unknown. Quarterly roadside surveys should be continued, and island-wide surveys should be conducted at 5-year intervals to assess changes in the population across the island. Finally, an island-wide conservation plan for the Tinian Monarch should be developed to help manage for the long-term conservation of the species.


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## Introduction

The Tinian Monarch (Monarcha takatsukasae), locally known as Chichirikan Tinian, is a small forest bird endemic to the island of Tinian in the Commonwealth of the Northern Mariana Islands (CNMI). First described by Takatsukasa and Yamashina (1931), this small (15 centimeters [6 inches]) songbird in the monarch flycatcher family (Monarchidae), has light rufous underparts, olive-brown upper parts, dark brown wings and tail, white wing bars, and a white rump and undertail coverts (Baker 1951). There are some reports that indicate the Tinian Monarch may once have occurred on the island of Saipan, approximately 5 km north of Tinian (Peters 1996). Reports of the species on the small island of Aguiguan, approximately 9 km south of Tinian, have generally been discounted (Engbring et al. 1986).

The Tinian Monarch was listed as endangered in 1970 under the Endangered Species Act (ESA) as its numbers were believed to be critically low because of removal of native forests for sugarcane production by the Japanese before World War II and destruction of forest as a result of this conflict (Baker 1946; USFWS 1970, 1987). Destruction of the native forest actually began in the $18^{\text {th }}$ century when the Spanish used Tinian as a supply island for Guam and introduced large herds of cattle and other ungulates to the island (Fosberg 1960). After World War II, it is believed that the U.S. Department of Defense (DOD) seeded the island with tangantangan (Leucaena leucocephala), a rapidly growing tree native to Mexico and Central America, in order to combat erosion (USFWS 1996). Based on reports of monarch numbers after the war, it is likely that while common in small patches of remaining forest, their distribution was highly restricted and their population small because so much of the forest had been destroyed (Downs 1946; Gleize 1945; Marshall 1949; USFWS 1987, 2005).

An island-wide survey of Tinian conducted in 1982 found the Tinian Monarch to be the second most abundant bird on the island with an estimated population of 39,338 birds widely distributed over the island in all forest types (Engbring et al. 1986). Based on these data, the species was downlisted from endangered to threatened in 1987 (USFWS 1987). A life history study conducted in 1994-1995 estimated the monarch's population to be approximately 52,904 birds based on densities in 3 forest types (USFWS 1996). A second island-wide survey of forest birds that used the same transects and methods as in 1982 estimated the species' population at 55,721 birds (Lusk et al. 2000). This latter study also found that vegetation density increased in all forest types since 1982, leading Lusk et al. (2000) to hypothesize that increases in monarch density were correlated with increases in vegetation density. Based on the finding that the Tinian Monarch had increased in number or was at least stable, that the primary listing factor (loss of forest habitat) had been ameliorated, and that the species was not currently threatened by other factors, the monarch was delisted in 2004 (USFWS 2004).

Other threats to the Tinian Monarch include predation by introduced mammals, random demographic fluctuations, disease, and the potential for the introduction of the brown treesnake (Boiga irregularis) to Tinian. The brown treesnake is responsible for the extinction and extirpation of 10 of the 13 native forest bird species on Guam (Conry 1988, Rodda et al. 1999, Savidge 1987). The brown treesnake is not established on Tinian but there have been at least seven snake sightings between 1994 and 2003 (Hawley 2002, USFWS 2004). None of the snakes were caught but descriptions of several of them were consistent with the brown treesnake (USFWS 2004). Brown treesnakes could potentially reach Tinian via cargo shipped from Guam,
where the snake is well established and occurs in extremely high densities (Rodda et al. 1999, USFWS 2004). It is hoped that measures to control brown treesnakes around ports and the inspection of high-risk cargo leaving Guam will prevent their accidental introduction to other islands. Regardless, the brown treesnakes is thought to in the process of becoming established on Saipan, though clear evidence of establishment or recruitment is lacking (Rodda and Savidge 2007, USFWS 2004).

Section $4(\mathrm{~g})(1)$ of the ESA requires the U.S. Fish and Wildlife Service (USFWS) to implement a system in cooperation with the States to effectively monitor (for no less than five years) the status of all species that have been delisted. The purpose of this post-delisting monitoring is to verify that the delisted species remains secure from the risk of extinction after removal from the protection of the ESA. In 2004, the USFWS published a final post-delisting monitoring plan (PDMP) for the Tinian Monarch (USFWS 2005).

This report summarizes the results of the Tinian Monarch post-delisting monitoring conducted between 2006 and 2011. It specifically outlines the results of the roadside surveys, small-scale early warning system plots, and forest cover monitoring identified in the PDMP. The results of island-wide surveys summarized in Camp et al. (2012) and U.S. Navy (2014) are also discussed in this report.

## Methods

## Roadside Surveys

Quarterly roadside surveys were conducted at 50 stations distributed around the island of Tinian by staff from the CNMI's Division of Fish and Wildlife (DFW) beginning in 1999 (Figure 1). All surveys were completed between one half-hour before sunrise and 10:30 a.m. and were only conducted under fair weather conditions when winds were judged to be less than three as measured on the Beaufort Scale. Trained observers recorded detections of all birds seen or heard within an unlimited distance from the station during a three-minute period at each station in January, April, July and October annually. Cloud cover, rain, and wind conditions were also recorded.

Trends in Tinian Monarch detections were assessed using generalized linear mixed models (GLMMs) with the lme4 (Bates et al. 2015) and glmmTMB (Brooks et al. 2017) packages in R (R Core Team 2020). Survey station and observer were fitted as random effect variables to account for lack of independence of counts from the same survey stations, not all stations being surveyed in all survey periods, and changes in observers over the survey period. Models were fitted assuming either a negative binomial or poisson distribution and were evaluated using AIC. Model assumptions were tested using residual plots, nonparametric dispersion test, and zeroinflation test from the DHARMa package (Hartig 2020).

## Small-Scale Early Warning System Plots

Three small-scale study plots, Santa Lourdes Shrine, the Airport Mitigation Area, and the Seaport Tangantangan Site (Figure 1), were established in Tinian Monarch habitat located near areas where brown trees snakes might arrive on Tinian, such as the airport, seaport, cargo offloading zone or staging areas, and/or sites that are the ultimate destination of materials brought from the other islands. As Tinian Monarch densities are generally higher in limestone forest
(Camp et al. 2012, USFWS 1996), we focused on establishing study sites in this habitat type. Two sites were established in limestone forest, Santa Lourdes Shrine and the Airport Mitigation Area. However, limestone forest habitat was limited near the Seaport, therefore, the Seaport Tanagantangan Site was established in mixed stand of tangantangan and secondary forest near the port. The Santa Lourdes site was located on the edge of the main village on the island, San Jose, in between the seaport and airport. The Airport Mitigation and Seaport Tangantangan Site were placed near the airport and seaport, respectively.

Tinian Monarchs at each study site were mist-netted, weighed, measured (wing, tarsus, tail, and bill length, and body mass), inspected for health, molt, and breeding condition. Each monarch was fitted with an aluminum band and a unique combination of three colored leg bands, then released unharmed in the same location where they were captured. Feather samples were collected from 61 individuals in 2008 and 2009 for genetic sexing. One individual was sampled twice resulting in 62 samples total. Thirty-five samples were analyzed to determine sex in 2011. Twenty-seven samples from 2008 were initially thought to be lost but were rediscovered in 2016. Two sets of methods, therefore, were employed to sex the samples genetically.


Figure 1. Location of small-scale early warning system plots and breeding bird survey stations on the island of Tinian.

Genomic DNA was extracted from 35 feather samples using the Qiagen Tissue Extraction kit, following the manufacturer's protocol, for the 2011 assessment. All individuals were independently sexed at least twice using protocols described by Fridolfsson and Ellegren (1999). PCR products were verified with agarose gel electrophoresis. The 2016 assessment utilized the
remaining 27 feather samples, which were stored at room temperature over 8 years and were considered degraded. As such, the 2016 assessment used the forward and reverse primer set Z37B, described in Dawson et al. (2015). PCR amplification of each sample followed the protocol outlined by Dawson et al. (2015). Genotyping was conducted by the Advanced Studies in Genomics, Proteomics and Bioinformatics facility at the University of Hawaii at Manoa using an Applied Biosystems 3730XL DNA Analyzer. Sex determination through peak visualization was conducted in the PeakScanner vers. 2 (Applied Biosystems).

Banded Tinian Monarchs were initially sexed using a combination of genetic sexing results and information on breeding status (e.g., presence of a brood patch) and breeding pair composition. Morphometric data (i.e., wing, tail, bill, and tarsus length and body mass) from these sexed individuals were then evaluated to determine if the morphometric measurements could be used to sex the remaining unsexed Tinian Monarchs. Overall differences in morphological measurements between the sexes were analyzed with a multivariate analysis of variance (MANOVA) with sex as the response and wing chord, bill length, long tarsus length, tail length, and weight as predictors. One-way Analysis of Variance (ANOVA) was used to check for significant differences in each individual morphometric. A linear discriminant function analysis (DFA) was used to identify the best morphometric characters to predict sex. Models with one, two, or all variables were evaluated. Model performance was based on accuracy scores derived from 5-fold repeat cross-validation. Prior to the MANOVA and DFA, outliers and potential morphological measurement errors were identified using box plots (univariate) and cook's distance (multi-variate). Correlation in morphometric variables was assessed using the Pearson's correlation coefficient and variables with a correlation coefficient of 0.75 or greater were dropped from the analysis. Normality and homogeneity of variance assumptions were tested with quantile-quartile plots, Shapiro-Wilk normality test, and Bartlett's test of homogeneity of variances. All analyses were completed in R using functions from the MASS (Venables and Ripley 2002) and caret (Kuhn 2020) packages.

Territories at each study site were mapped annually by recording the location of all marked and unmarked Tinian Monarchs and their nest sites using general territory mapping methods outlined in Bibby et al. (2000). Whenever possible, territorial and breeding behaviors were noted to delineate territorial boundaries and each site was mapped for at least 5 days by two or more observers. These maps were then used, with the data on marked birds, to estimate overall territory occupancy (i.e., the likelihood a territory would be occupied each year) and territory turnover rates (i.e., the likelihood a banded bird in a territory will no longer occur in a territory the subsequent year) by each sex. Territory occupancy was estimated by comparing territory maps for each study plot for each monitoring year and scoring a territory as occupied or unoccupied each monitoring year. These scores were then used to estimate the likelihood of a territory being occupied from year to year using the island R package (Ontiveros et al. 2019). We assumed all territories were mapped accurately, independent, and equivalent and that the birds occupying each territory were identified correctly. Likelihood of territory turnover for each sex was estimated by comparing territories occupied by a known-sex marked individual from one year to the next. Turnover was calculated as the percentage of territories in one year not occupied by the same individual in the subsequent year. Like the estimate of territory occupancy, we assumed all birds occupying each territory were identified correctly. We examined variation in territory size using a general linear model, with year and site as factors.

We estimated apparent annual survival ( $\varphi$ ) and encounter probability ( $\rho$ ) for banded Tinian monarchs between 2006 and 2010 using Cormack-Jolly-Seber models. We reported apparent annual survival as the population size was unknown and the true fate of marked birds was unknown. Modeling parameters initially included sex and study area as we expected survival and encounter probability could vary between these variables (see Table 5 for a list of the final models evaluated). A goodness of fit test was carried out on a global model ( $\varphi s e x+$ site. $\rho s i t e$. using the median $\mathrm{c}^{\wedge}$ approach. The dispersion parameter indicated moderate over-dispersion ( $\mathrm{c}^{\wedge}=$ 1.41). Therefore, we adjusted the value of c and conducted model selection using the resulting QAICc values, as recommended by Burnham and Andersen (2002). We considered the model with the lowest QAICc to have the best fit, but also considered models with QAICc values that differed ( $\Delta \mathrm{QAICc}$ ) by $\leq 2.0$. All analyses were conducted in the statistical program R. Model fitting and testing was conducted using the Marked package (Laake et al. 2013). The mra (McDonald 2018) package was used for global model goodness of fit testing and calculating c ${ }^{\text {. }}$.

Monitoring was conducted at small-scale plots primarily between 2006 and 2010. The Santa Lourdes and Airport Mitigation sites were established and monitored in 2006 while the Seaport Tangantangan site was established in 2007. Monitoring was conducted at all three sites in 2007, 2008, 2009, and 2010. The monitoring in 2010 was originally expected to be the final year of monitoring. However, some funds were obtained to conduct a fifth year of monitoring at the Seaport Tangantangan site in 2011. Therefore, the results for these estimates focus on the 2006 to 2010 monitoring period for the Santa Lourdes and Airport Mitigation sites and 2007 to 2011 for the Seaport Tangantangan site.

## Land-use Monitoring

Trends in land use were estimated over the monitoring period using two methods. First, landcover estimates for the island of Tinian were obtained from the Forest Service (Liu and Fisher 2006). These estimates were based on a 2004-2005 mapping assessment using Quickbird and Ikonos satellite imagery of the island and were considered the baseline estimates for habitat availability on the island at the start of the monitoring period. Forest clearings were then identified and mapped on 2010 Worldview 2 satellite imagery of Tinian. If these clearings occurred in areas that were classified as forest in 2005 they were considered forest clearings and the total acreage of forest cleared was estimated for the monitoring period. The second monitoring method involved tabulating applications for habitat clearing to the Division of Fish and Wildlife (DFW) during the monitoring period. All land clearing in the CNMI requires a permit by the Division of Environmental Quality, which includes clearance forms from DFW.

## Results

## Roadside Surveys

A total of 16 quarterly surveys were conducted on Tinian between November 1999 and July 2010 by CNMI DFW. No surveys were completed during 2001 to 2003 and 2006 to 2009 resulting in large gaps in the data set (Figure 2). The best fitting model for Tinian Monarch detections was the zero inflated negative binomial (ZINB) model with observer as a random effect (Table 1). Tinian Monarchs showed a significant decline in detections between 1999 and 2010 (Table 2, Figure 2).

Table 1. Models evaluated to assess changes in Tinian Monarch roadside counts from 1999 to 2010.

| Model | Distribution ${ }^{\mathbf{1}}$ | AIC |  |
| :--- | :--- | :--- | :--- |
| TAIC | Survey $+(1 \mid$ Observer $)$ | ZINB | 2028.3 |
| TIMO $\sim$ Survey $+(1 \mid$ Observer $)$ | NB | 0 |  |
| TIMO $\sim$ Survey $+(1 \mid$ Station $)$ | ZINB | 2041.0 | 12.7 |
| TIMO $\sim$ Survey | NB | 2049.2 | 20.9 |
| TIMO $\sim$ Survey $+(1 \mid$ Station $)$ | NB | 2059.2 | 20.9 |
| TIMO $\sim$ Survey $+(1 \mid$ Observer $)$ | PO | 2352.2 | 21.9 |
| TIMO $\sim$ Survey $+(1 \mid$ Station $)$ | PO | 324.4 |  |
| TIMO $\sim$ Survey | PO | 2369.4 | 341.1 |

${ }^{1}$ Zero inflated Negative Binomial (ZINB), Negative Binomial (NB), Poisson (PO).
Table 2. Posterior mean values, standard errors, and $95 \%$ credible intervals for the parameters in the zero inflated negative binomial mixture model for Tinian Monarch roadside survey detections.

| Coefficient | Mean | Lower 95\% | Upper 95\% |
| :--- | :---: | :---: | :---: |
| Intercept | 1.003 | 0.460 | 1.547 |
| Survey Sequence | -0.020 | -0.038 | -0.001 |
| Observer (intercept) | 0.436 | 0.242 | 0.787 |
| Zero Inflation (intercept) | -0.322 | -0.670 | 0.027 |



Figure 2. Modelled trend, with 95\% confidence interval, of Tinian Monarch detections at roadside stations surveyed from 1999 to 2010. Black diamonds indicate actual survey periods.

## Small-Scale Early Warning System Plots

A total of 119 Tinian Monarchs were banded during the 2006-2011 monitoring period; 39 at Santa Lourdes, 51 at the Airport Mitigation, and 30 at the Seaport Tangantangan site (Appendix A). One hundred and five (105) were after hatch year birds, 10 were hatch year birds, and 4 were nestlings. Sexes were identified for 53 after hatch year individuals. One of these individuals (color combination ARRG) had incomplete morphometric measurements and was dropped from the analysis. An additional eight individuals (color combinations AKKP, ARWB, BAWG, BRGA, KBAP, GAGR, RWKA, and WRAG) were found to have extreme or inconsistent measurements, potentially due to measurement or data transcription errors, based on univariate, bivariate, and multivariate assessments. These individuals were also dropped from the morphological assessment resulting in a total sample of 44 birds ( 25 females and 19 males) available for analysis.

The morphological measurements of the 44 sexed birds were not significantly correlated and all measurements met the assumptions of normality and homogeneity of variance. A multivariate analysis of variance (MANOVA) on the 44 birds showed a significant morphological difference between females and males $\left(\mathrm{F}_{1,38}=7.369, \mathrm{p}<0.001\right)$. All five morphological measurements overlapped between the sexes (Table 3). Of the five measurements evaluated, wing chord was significantly longer for males than females (Table 3).

Table 3. Morphological differences between male and female adult Tinian Monarchs. F values are derived from univariate ANOVA.

| Measurement | Sex | n | Mean | SD | F value, df | P value |
| :--- | :--- | :---: | :---: | :---: | :--- | :--- |
| Wing chord $(\mathrm{mm})$ | Female | 25 | 66.6 | 1.58 | 15.471 | 0.0003 |
|  | Male | 19 | 68.7 | 1.99 | 1,42 |  |
| Tail length $(\mathrm{mm})$ | Female | 25 | 66.3 | 1.73 | 1.068 | 0.31 |
|  | Male | 19 | 66.9 | 2.04 | 1,42 |  |
| Bill length $(\mathrm{mm})$ | Female | 25 | 14.1 | 0.77 | 1.526 | 0.22 |
|  | Male | 19 | 14.3 | 0.83 | 1,42 |  |
| Tarsus length $(\mathrm{mm})$ | Female | 25 | 22.9 | 0.73 | 1.239 | 0.27 |
|  | Male | 19 | 22.6 | 0.82 | 1,42 |  |
| Weight $(\mathrm{g})$ | Female | 25 | 11.8 | 0.92 | 1.297 | 0.26 |
|  | Male | 19 | 11.5 | 0.98 | 1,42 |  |

Linear DFA models were developed using the morphological data from 44 known sexed individuals. Table 4 summarizes the accuracy assessments of the various DFA models built to discriminate sex using the five morphological variables. The top performing model, based on accuracy was the model which included wing and tarsus measurements (Table 4). The resulting discriminant function ( 0.657 (wing chord) -1.073 [tarsus length]) had an overall testing accuracy of $78 \%$ and a $74 \%$ and $80 \%$ testing accuracy for males and females, respectively. The model with just wing ( 0.530 [wing chord]) had an overall testing accuracy of $73 \%$ and a $68 \%$ and $77 \%$ testing accuracy for males and females, respectively. These two models were used to help determine the sex of the 53 unsexed adult Tinian Monarchs for the following survival analysis.

Table 4. Linear discriminant function analysis models for predicting sex of adult Tinian Monarchs based on morphometric measurements. Accuracy scores are averages based on repeated 5 -fold cross-validation.

| Model | Testing Accuracy <br> $( \pm \mathbf{S D})$ | Testing Kappa <br> $( \pm \mathbf{S D})$ |
| :--- | :---: | :---: |
| Wing + Tarsus | $78 \pm 20$ | $0.54 \pm 0.41$ |
| Wing + Tail + Bill + Tarsus + Weight | $77 \pm 18$ | $0.52 \pm 0.40$ |
| Wing + Weight | $73 \pm 20$ | $0.44 \pm 0.43$ |
| Wing | $73 \pm 20$ | $0.45 \pm 0.39$ |
| Wing + Tail | $72 \pm 24$ | $0.42 \pm 0.49$ |
| Wing + Bill | $69 \pm 20$ | $0.37 \pm 0.41$ |
| Bill | $60 \pm 18$ | $0.18 \pm 0.38$ |
| Tarsus | $57 \pm 20$ | $0.06 \pm 0.40$ |
| Tail | $57 \pm 18$ | $0.07 \pm 0.38$ |
| Weight | $55 \pm 16$ | $0.01 \pm 0.32$ |

Resight data were available on 104 Tinian monarch at the three study sites ( 32 at Santa Lourdes, 46 at the Airport Mitigation, and 26 at the Seaport Tangantangan; Appendix B). Due to the limited sample sizes, we only evaluated a limited number of mark-recapture models (Table 5). The model with all parameters constant performed the best (Table 5). Although the model with sex was within $2 \Delta \mathrm{QAICc}$, it had little effect on model deviance and was considered an uninformative parameter as per Arnold (2010). Based on the top model, estimated annual survival was $78 \%(\mathrm{SE}=3 \%, 95 \% \mathrm{CI}=72-83 \%)$ and encounter probability was $95 \%(\mathrm{SE}=2 \%$, $95 \% \mathrm{CI}=88-98 \%$ ).

Table 5. Models used to investigate apparent annual survival $(\varphi)$ and encounter probability ( $\rho$ ) for Tinian monarchs between 2006 and 2010 on the island of Tinian, Mariana Islands, USA. Subscripts indicate whether parameters differed among groups (e.g., $\varphi$ sex) or were constant ( $\varphi$ ). $\Delta$ QAIC $_{c}$ is the difference from the best model, which had a QAIC $_{c}=188.16, w^{2}$ QAIC $_{c}$ is the $\mathrm{QAIC}_{\mathrm{c}}$ weight, and QDev is the Quasi Deviance.

| Model | Number of <br> parameters | $\Delta$ QAIC $_{\mathbf{c}}$ | wQAIC $_{\mathbf{c}}$ | QDev |
| :--- | :---: | :---: | :---: | :---: |
| $\varphi . \rho$. | 2 | 0.00 | 0.525 | 369.57 |
| $\varphi$ sex $\rho$. | 3 | 1.37 | 0.264 | 368.06 |
| $\varphi_{\text {time }} \rho$. | 5 | 3.22 | 0.105 | 362.99 |
| $\varphi$ site $\rho$. | 4 | 3.97 | 0.072 | 368.94 |
| $\varphi_{\text {sex }+ \text { site }} \rho$. | 5 | 5.52 | 0.033 | 367.62 |

Territory size was larger in the Seaport Tangantangan site ( $0.42 \pm 0.03 \mathrm{ha}$ ) than in the Airport Mitigation and Santa Lourdes Shrine sites ( $0.12 \pm 0.005$ ha and $0.13 \pm 0.006$ ha, respectively; $\left(\mathrm{F}_{2,152}=147.32, \mathrm{p}<0.001\right.$ ). Territory size did not vary among years so data from all three years were combined.

Data to estimate annual overall territory occupancy and turnover rates by sex were available for the Santa Lourdes and Airport Mitigation sites for the 2006 to 2007, 2007 to 2008, 2008 to 2009,
and 2009 to 2010 periods (see Appendix C for annual territory maps). Territory data for the Seaport Tangantangan site were only available for 2008 to 2009, 2009 to 2010, and 2010 to 2011. Monitoring was initiated at this site in 2007. However, only a small area was sampled due to time constraints.

A total of 1.7, 1.2, and 2.7 hectares of forest were consistently sampled at the Airport Mitigation, Santa Lourdes, and Seaport Tangantangan sites, respectively, to measure territory occupancy and turnover between years. The overall annual territory occupancy rate for all study sites combined was $94 \%$ ( $95 \%$ CI: 91-97\%). Annual territory occupancy for the Airport Mitigation, Santa Lourdes, and Seaport Tangantangan sites was 96\% (95\% CI: 94-99\%), 92\% (95\% CI: 88-97\%), and $90 \%$ ( $95 \%$ CI: $89-97 \%$, respectively. Finally, average annual territory turnover rates for males and females over all study sites was $35 \%$ ( $95 \%$ CI: $24-46 \%$ ) and $42 \%$ ( $95 \%$ CI: $31-53 \%$ ), respectively.

An unexpected result of the netting in 2006 was that 15 of the monarchs (39\%) had scabby lesions on their feet and toes typical of those caused by avian pox virus (Poxvirus avium, Figure 3 ). Another seven monarchs ( $18 \%$ ) had missing or deformed toes typical of healed pox lesions. These pox-like lesions were also observed on monarchs in 2007 and 2009 but not in 2008. In 2007, the prevalence of the lesions ( $11 \%$ ) was lower than in $2006\left(39 \% ; \mathrm{X}^{2}=8.49, \mathrm{df}=1, \mathrm{p}=\right.$ 0.004 ). Lesions were only observed on one monarch in 2009 (3\%). This lower occurrence along with the lack of observations of pox-like lesions in 2008 may indicate that we observed the tail end of an epizootic or that it may occur in cycles. In 2006 there was some indication that prevalence of these lesions was higher at the Santa Lourdes Shrine (44\%) than at the Airport Mitigation Area (36\%). In 2007 prevalence was again higher at Santa Lourdes (19\%) than at the Airport Mitigation ( $0 \%$ ) and was also high at the Seaport Tangantangan site (20\%). The Santa Lourdes Shrine and the Seaport sites are closer to urbanized areas of Tinian where mosquito abundance may be higher due to the presence of standing water in residential structures and abandoned machinery that provide mosquito breeding sites. These sites are also closer to populations of domestic birds such as chickens, which could serve as a reservoir for disease. Strangely, the pox-like lesions were observed only in monarchs and not in any of the other species captured.


Figure 3. Photograph of pox-like lesions on a Tinian Monarch's foot. Photograph by Eric VanderWerf.

Tissue samples were collected in 2006 from lesions on seven monarchs to conduct clinical tests. These samples were collected from monarchs only if there was a loose scab that could be easily removed without injuring the bird or exacerbating the lesion. Drs. Dennis Triglia and Carter Atkinson of the U.S. Geological Service Biological Resources Discipline in Hawaii attempted to test whether those lesions were caused by avian pox virus using tissue culture techniques. They were able to amplify and sequence Avipoxvirus ( 4 b core protein) from one of the cell culture pellets from 2006. A check on Genbank indicates that it is $100 \%$ identical to Canarypox that has been reported from both Hawaii and the Galapagos Islands.

## Land-use Monitoring

Approximately 165 hectares of forest were cleared on Tinian between 2005-2006 and 2010 based on an assessment of satellite imagery. Sixty-nine percent (114 ha) of this total was potential Tinian Monarch habitat (tangantangan thicket and mixed introduced forest) while the majority of the remainder was shrub and grass areas. If we assume that the clearings occurred consistently over a 5-year period, this would be equivalent to 22.8 ha of Tinian Monarch habitat cleared per year or 33 ha of land cleared per year.

Data on 79 land clearing permits were obtained from DFW for 2007 thru 2010 (DFW, unpubl. data). The area expected to be cleared was not reported for eight of the permits, so those permits were dropped from the analysis. The total acreage cleared under the remaining permits was 78 hectares, with the majority ( $68 \%$ ) of the permits being for small areas ( $<1 \mathrm{ha}$ ). The largest clearing was 20 ha in Kahet Acra in 2007. Based on standard error, the average clearing size was similar between years (Figure 4) and average total area cleared per year was 19.5 ha ( $\pm 8.6 \mathrm{SD}$ ).


Figure 4. Average clearing size per year ( $\pm$ SE) based on permits for clearing to the Division of Fish and Wildlife in 2007-2010.

## Discussion

Results from the roadside surveys showed a significant decline in counts of Tinian Monarchs from 2001 to 2010. This is consistent with the reported island-wide population declines of the species from 1996 to 2008 by Camp et al. (2012). However, it is uncertain if counts remained low between 2010 and 2013, when a subsequent island-wide survey indicated the decline in the island-wide population had reversed to pre-2008 levels (U.S. Navy 2014). The 2013 and earlier island-wide surveys appear to indicate that the Tinian Monarch population may be cycling through time. However, regular quarterly roadside surveys and more frequent island-wide surveys are needed to determine if this is the case (see Recommendations below).

Tinian Monarch annual survival rates ( $78 \%, 95 \% \mathrm{CI}=72-83 \%$ ) from the small-scale warning plots fell within the lower part of the range of survival rates reported for other Pacific Island monarch flycatchers. Blanvillain et al. (2003) reported annual survival rates ranging from 8592\% for the Tahiti Flycatcher (Pomarea nigra). Robertson et al. (1994) reported annual survival rates of $76 \%$ for the Rarotonga Flycatcher (Pomarea dimidiate) prior to rat control. VanderWerf (2008) reported male and female annual survival rates of $87 \%$ and $81 \%$, respectively, for the Hawai` ` `lepaio (Chasiempis sandwichensis), and $84 \%$ ( $86 \%$ with rodent control) and $55 \%$ ( $82 \%$ with rodent control) for male and female O'ahu 'Elepaio (VanderWerf 2009).

We were unable to evaluate potential differences in annual survival by sex in this study due to limited data. Average annual territory turnover rates, however, were higher for females (42\%) than males ( $35 \%$ ) and studies by Greenberg and Gradwohl (1986) and Woltmann and Sherry (2011) on Central American insectivorous birds showed a consistent relationship between territory turnover and mortality. Additional research on survival rates of male and female Tinian Monarchs may also show a difference between sexes.

Despite having relatively low annual survival for Pacific Island monarch flycatchers, the overall territory occupancy rate was high ( $94 \%$ ). No data were collected on nesting success and juvenile survival in this study. However, the high territory occupancy rate may indicate good levels of recruitment. Additional data are needed to confirm this.

Territory size estimates for the Airport Mitigation site were remarkably similar to the average territory size determined for that area in 1994 to 1995 , which was $0.12 \pm 0.02 \mathrm{ha}(\mathrm{n}=17)$, indicating that the monarch territory sizes remained stable over this period (USFWS 1996). Territory size estimates for the Airport Mitigation site and Santa Lourdes site were also similar, indicating potentially similar estimates for native forest dominated sites. Territory size at the Seaport Tangantangan site, a mixed tangantangan and secondary forest site, was smaller than the territory sizes reported at tangantangan $(0.64 \pm 0.08 \mathrm{ha}(\mathrm{n}=9))$ and secondary forest $(0.51 \pm 0.06$ ha $(\mathrm{n}=8)$ ) sites sampled in 1995. However, it fell within the range of territory sizes (0.35-0.58 ha) recorded at two tangantangan and two secondary forest territory mapping plots sampled in August 2008 (USFWS 2009).

Comparison with densities derived from point-transect surveys indicate much higher densities in the small-scale plots (Camp et al. 2012, U.S. Navy 2014). For example, Camp et al. (2012) reported densities of $6.5,5.5$, and 4.0 birds per hectare in limestone forest, secondary forest, and tangantangan habitats. These differences are likely due to the areas sampled using each method.

The small-scale plots were limited to three areas of generally contiguous habitat while the pointtransect stations generally sample a mosaic of habitats. Sampling small-scale plots over a wider range of areas on Tinian would likely show a broader range of potential densities.

The differences in territory sizes and monarch densities among native dominated, secondary forest, and tangantangan sites may be related to plant diversity. Habitat sampling along bird survey transects on Tinian by Vogt (2009) showed a positive correlation between Tinian monarch detections and tree diversity and native forest sites generally have greater plant diversity than secondary forest and tangantangan sites (Fosberg 1960, Perry and Morton 1999). Higher plant diversity may increase the number of potential foraging and nesting sites available and may have an impact on prey availability. However, further work is needed to confirm this.

As noted in the Results, an unexpected outcome of the monitoring was the observation and identification of avian pox in Tinian Monarchs. Avian pox virus is found worldwide and has been reported in domestic fowl on Guam (Savidge et al. 1992). To our knowledge, however, it has not been reported previously in wild birds in the Mariana Islands. Avian pox is fairly common in Hawaii and is thought to be one of the main threats to many Hawaiian forest birds (Atkinson et al. 2005, VanderWerf et al. 2006, van Riper et al. 2002). Pox has also been recently identified in the Galapagos Islands (Thiel et al. 2005) where it may have been introduced in the 1890s (Parker et al. 2011). Immunity to pox varies substantially among bird species, and many continental species are not seriously affected by the virus. The degree of immunity to pox in birds from the Mariana Islands is unknown. However, we only observed pox-like lesions in monarchs and not in any of the other species captured during the study. Our results also showed a decline in prevalence from a high of $38 \%$ in 2006 to $0 \%$ and $3 \%$ in 2008 and 2009, respectively, indicating we observed the tail end of an epizootic or that it may occur in cycles.

## Recommendations

Based on the annual monitoring and this summary assessment we have three sets of recommendations:

1. Continue Long-term Monitoring - Tracking changes in Tinian Monarch populations and their habitats is important for ensuring that changes in the population are identified and adequately addressed prior to implementing legal changes to the status of the species. Several monitoring methods were identified in the PDMP, the following three methods are recommended for longer-term monitoring.
a. Continue to conduct roadside surveys quarterly - Historically, the roadside surveys were not consistently implemented due to lack of personnel and other factors. If possible, these surveys should be conducted quarterly with trained personnel to allow for tracking changes in Tinian Monarch detections. An analysis of the full survey dataset should additionally be conducted similar to the work by Ha et al. (2018) for the roadside surveys on Saipan.
b. Repeat island-wide surveys at 5-year intervals - Surveys were conducted in 2008 and 2013. The 2008 survey indicated that Tinian Monarch densities had declined both temporally and spatially (Camp et al. 2012) while the 2013 survey indicated the population had returned to levels similar to the surveys prior to 2008 (U.S. Navy 2014). As the Tinian Monarch population appears to fluctuate it's
recommended that island-wide surveys be conducted at $5-\mathrm{yr}$ intervals for 20 years to better track the population.
c. Continue land clearing permit monitoring and monitor long-term ( 10 years) landcover changes using satellite imagery - Tracking annual land clearing permits and mapping landcover changes using satellite imagery was found to be a useful tool for assessing changes in Tinian Monarch habitats. Though these methods showed little change in the land-use and land-cover on Tinian during the monitoring period, they will be important for tracking changes associated with future projects. For example, several development projects have been proposed and implemented on Tinian since the monitoring period completed in 2011. In 2015 the Department of Defense released a draft environmental impact statement for a CNMI Joint Military Training (CJMT) project. This project proposed to establish a series of live-fire and maneuver ranges and training areas on military leased lands on Tinian, including in or near the Airport Mitigation study plot (U.S. Navy 2015). Areas have been cleared since 2014 for the West San Jose Homestead project and since 2016 for the Altercity Resort (formerly the Plumeria Resort) project, both near the Seaport Tangantangan study plot (U.S. Navy 2015, Willsey in litt. 2016). A Water Transfer Station project was also begun in 2017 in the non-leased area of the island (U.S. Navy 2015).

Data on land clearing permits should continue to be collected and evaluated regularly to identify spatial trends in development and their relationship to Tinian Monarch habitats. Large-scale land-cover changes should also be evaluated at approximately 10 -year intervals using satellite imagery. The NOAA Coastal Change Analysis Program (CCAP) regularly maps landcover in the Mariana Islands, which can be used to track landcover changes spatially. These data can then be tied into roadside and island-wide survey results to evaluate spatial trends in Tinian Monarch populations.
2. Develop island-wide Tinian Monarch conservation strategy - Areas needed to ensure the long-term survival of the Tinian Monarch on Tinian have not been identified and, currently, only the Federal Aviation Administration Mitigation Area had been set aside for Tinian Monarch conservation, though even this area is being proposed for clearing as part of a live range (see U.S. Navy 2015). Habitat loss was the main threat that lead to the original listing of this species (USFWS 2004). Evidence from this study and previous research (USFWS 1996) suggests that densities of Tinian Monarchs are highest in native limestone forest. Protecting the remaining forest and restoring other areas may increase the distribution and density of the Tinian Monarch in areas currently not used or with low numbers of birds.

An island-wide conservation plan would be a helpful tool to identify these areas and help ensure that they be considered in development projects planned for the island. This strategy should be developed in conjunction will all relevant partners (e.g., Tinian Mayor, Tinian DLNR, CNMI DLNR, and U.S. Navy) and should consider Tinian Monarchs throughout the island.
3. Conduct additional research - Information about the biology and conservation of the Tinian Monarch is largely limited to data collected in this and the mid-90s study. Currently, there is still not much known about the Tinian Monarch's population dynamics and threats. The following two areas need additional research:
a. Population Biology - Demographic based Population Viability Assessments (PVAs) are important tools for conservation planning and the Tinian Monarch conservation strategy identified earlier would benefit from this type of assessment. However, data on the population biology of the Tinian Monarch are still needed before these assessments can be developed. For example, overall annual survival rates for adults have been estimated, juvenile survival has not been estimated and adult survival rates have not been estimated by habitat or sex. Some data on nesting success were collected during the mid-1990s. However, nesting success data for all habitats is not available. A demographic study, like the Tropical Monitoring Avian Productivity and Survivorship (TMAPS) project on Saipan (Saracco et al. 2014), that collects these data in all forested habitats is needed.
b. Avian Pox - Though this study identified that avian pox occurs in the Tinian Monarch population, its prevalence and impact on the species is unknown. Research on avian poxvirus and Tinian Monarchs is needed to address its potential impact on the species. This study could be conducted in conjunction with the Population Biology study identified previously.

Avian poxvirus is transmitted primarily by mosquitoes in Hawaii (van Riper et al. 2002, VanderWerf et al. 2006), but it can also be transmitted by physical contact with another bird or an infected surface, such as a branch or nest (Tripathy 1993). In order to prevent potential spread of disease among birds captured during field work we recommend that all equipment be sterilized with a $10 \%$ bleach solution after any bird with lesions is captured, including banding pliers, wing ruler, calipers, bird holding bags, and researcher's hands. We also believe it would be useful to sample mosquito populations in different areas of the island to determine where the disease may be coming from. If mosquito breeding sites are located the mosquitoes should be eradicated by removing the sites or by treatment with mosquito larvicides.

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## Appendices

Appendix A - Banding records of Tinian Monarchs caught between 2006 and 2010.
Appendix B - Re-sight records of Tinian Monarchs banded between 2006 and 2010.
Appendix C - Study plot and Tinian Monarch territory mapping figures from the 2007-2010 annual monitoring report
Appendix A: Banding records of Tinian Monarchs caught between 2006 and 2010 during post-delisting monitoring work for the Tinian Monarch. Color codes: $\mathrm{AL}=$ Aluminum; $\mathrm{BL}=\mathrm{Blue} ; \mathrm{RD}=\mathrm{Red} ; \mathrm{GR}=$ Green; $\mathrm{PK}=\mathrm{Pink} ; \mathrm{BK}=\mathrm{Black} ; \mathrm{WH}=\mathrm{White}$. Age codes: $\mathrm{AHY}=$ After hatch year: $\mathrm{HY}=$ Hatch year: $\mathrm{L}=$ Local (not fledged). Sex codes: FB = Female with Brood Patch, FG $=$ Female based on Genetics, $\mathrm{FM}=$ Female based on Mate, $\mathrm{MG}=$ Male based on Genetics, $\mathrm{MM}=$ Male based on Mate, $\mathrm{U}=\mathrm{Unknown}$. Pox codes: $\mathrm{H}=$ Healthy; $\mathrm{A}=$ Active; $\mathrm{I}=$ Inactive or healed.

| Year | Month | Day | Location | Net | Capture | Prefix | Suffix | Leftleg | Rightleg | Age | Sex | Molt | Wing | Tail | Bill <br> Length | Tarsus long | Tarsus short | Wt bird | Pox <br> status |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 4 | 14 | Santa Lourdes | 1 | I | 2460 | 701 | AL/BL | BL/RD | AHY | U | Y | 71.5 | 71 | 15.6 | 23.9 | 23 | 13.4 | I |
| 2006 | 4 | 14 | Santa Lourdes | 2 | I | 2460 | 702 | AL/RD | GR/BL | AHY |  | N | 70 | 69 | 15 | 22.5 | 22.3 | 11.3 | A |
| 2006 | 4 | 14 | Santa Lourdes | 2 | I | 2460 | 703 | AL/GR | RD/RD | AHY | FB | N | 67.5 | 68 | 14 | 22.3 |  | 10.6 | I |
| 2006 | 4 | 14 | Santa Lourdes | 1 | I | 2460 | 705 | AL/PK | GR/GR | AHY | U | N | 66 | 66 | 14.2 | 22.4 |  | 10.4 | A |
| 2006 | 4 | 14 | Santa Lourdes | 2 | R | 2460 | 703 | AL/GR | RD/RD |  |  |  |  |  |  |  |  |  |  |
| 2006 | 4 | 14 | Santa Lourdes | 4 | I | 2460 | 706 | BL/AL | PK/PK | AHY | U | Y | 68 | 68 | 14.9 | 23.2 |  | 12.6 | H |
| 2006 | 4 | 15 | Santa Lourdes | 3 | I | 2460 | 709 | GR/AL | RD/BL | AHY | U | Y | 70.5 | 70 | 15 | 21.9 |  | 11.7 | H |
| 2006 | 4 | 15 | Santa Lourdes | 5 | I | 2460 | 710 | RD/PK | AL/GR | AHY | MM | Y | 70 | 69 | 15.2 | 23.6 |  | 11.8 | A |
| 2006 | 4 | 15 | Santa Lourdes | 5 | I | 2460 | 711 | BL/GR | AL/PK | AHY | FB | Y | 65.5 | 67.5 | 14.5 | 22.7 |  | 11.9 | A |
| 2006 | 4 | 15 | Santa Lourdes | 1 | R | 2460 | 701 | AL/BL | BL/RD |  |  |  |  |  |  |  |  |  |  |
| 2006 | 4 | 15 | Santa Lourdes | 1 | I | 2460 | 714 | GR/BL | AL/RD | HY | U | N | 66.5 | 69.5 | 15.1 | 23 |  | 12.3 | A |
| 2006 | 4 | 16 | Airport Mitigation | 2 | I | 2460 | 716 | GR/RD | AL/GR | AHY | U | N | 72 | 72 | 14.9 | 23.2 |  | 11.7 | I |
| 2006 | 4 | 16 | Airport Mitigation | 2 | I | 2460 | 717 | PK/GR | AL/BL | AHY | FB | N | 67 | 67 | 14.9 | 22.7 |  | 13.5 | H |
| 2006 | 4 | 16 | Airport Mitigation | 1 | I | 2460 | 718 | BL/RD | AL/PK | AHY | U | N | 72 | 69.5 | 15.8 | 24 |  | 12.6 | I |
| 2006 | 4 | 16 | Airport Mitigation | 3 | I | 2460 | 719 | GR/RD | AL/BL | AHY | U | N | 70 | 64.5 | 15.5 | 22.6 |  | 11.9 | H |
| 2006 | 4 | 16 | Airport Mitigation | 1 | I | 2460 | 723 | BL/AL | BK/RD | AHY | U | Y | 69 | 66.5 | 15.1 | 23 |  | 12.9 | H |
| 2006 | 4 | 16 | Airport Mitigation | 5 | I | 2460 | 725 | AL/GR | BL/GR | AHY | U | N | 71.5 | 71 | 15.7 | 24.4 |  | 13 | H |
| 2006 | 4 | 16 | Airport Mitigation | 5 | I | 2460 | 726 | AL/BK | GR/RD | HY | U | N | 63.5 | 65.5 | 15 | 23.6 |  | 11.6 | H |
| 2006 | 4 | 16 | Airport Mitigation | 6 | I | 2460 | 727 | AL/BL | PK/BL | AHY | U | Y | 69.5 | 69 | 15 | 22.6 |  | 13.5 | H |
| 2006 | 4 | 17 | Airport Mitigation | 6 | I | 2460 | 732 | RD/AL | PK/GR | AHY | U | N | 65 | 64.5 | 14.6 | 23 |  | 11.1 | H |
| 2006 | 4 | 17 | Airport Mitigation | 8 | I | 2460 | 733 | PK/AL | BL/BK | AHY | MM | Y | 71 | 70 | 15.5 | 23.5 |  | 12.6 | A |
| 2006 | 4 | 17 | Airport Mitigation | 7 | I | 2460 | 734 | GR/BK | AL/PK | AHY | FG | Y | 68 | 67.5 | 14.9 | 23.5 |  | 12.2 | H |
| 2006 | 4 | 17 | Airport Mitigation | 7 | I | 2460 | 735 | AL/RD | BL/RD | AHY | U | N | 68.5 | 70 | 15.2 | 22.2 |  | 12.1 | H |
| 2006 | 4 | 17 | Airport Mitigation | 6 | I | 2460 | 736 | BL/GR | AL/BK | AHY | U | Y | 65.5 |  | 14.7 | 23.4 |  | 12.8 | H |
| 2006 | 4 | 17 | Airport Mitigation | 4A | 1 | 2460 | 737 | BK/AL | GR/PK | AHY | MM | N | 69.5 | 67.5 | 15 | 23.3 |  | 13 | I |
| 2006 | 4 | 17 | Airport Mitigation | 4B | I | 2460 | 738 | GR/PK | PK/AL | AHY | U | N | 66.5 | 69 | 15.3 | 22.3 |  | 11.9 | A |


| \％ |  | 4 | － | $-$ | I | J | く |  | ＜ | ＜ | 4 | 7 | ＜ |  | ＜ | コ | ＜ | － |  | I |  | I | $\pm$ | － |  |  |  |  |  | ＜ | $\pm$ |  |
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| \％들 | $\begin{array}{\|c\|} \hline 0 \\ \mathrm{I} \end{array}$ | $\stackrel{ \pm}{=}$ | $\begin{array}{\|c\|} \underset{\mathrm{I}}{ } \\ \hline \end{array}$ | $\stackrel{\grave{\mathrm{j}}}{ }$ | $\vec{m} \mid$ | $\stackrel{ \pm}{=}$ | $9$ |  | $\begin{gathered} m \\ \underset{\sim}{n} \end{gathered}$ | $\underset{\mathrm{i}}{ }$ | $\begin{array}{\|c\|} \hline \underset{\text { In }}{ } \\ \hline \end{array}$ | $\stackrel{+}{\ddagger}$ | $\bigcirc$ |  | $\stackrel{\text { ¢ }}{ \pm}$ | $\underset{\sim}{n}$ | $=$ | $\stackrel{n}{=}$ |  |  |  |  | $\sim$ | $\sim$ |  |  |  |  |  |  | $\stackrel{+}{\square}$ |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\underset{\sim}{n}$ | $\stackrel{\infty}{\stackrel{\infty}{\lambda}} \mid$ |  |  |  | $\begin{aligned} & \infty \\ & \underset{\sim}{4} \end{aligned}$ | $\overrightarrow{~ N}$ | $\underset{\sim}{n}$ |  |  |  |  |  |  | $\stackrel{\wedge}{i}$ |  |
|  | $\left\|\begin{array}{c} \hat{X} \end{array}\right\|$ | $\underset{\sim}{c} \mid$ | $\cdots$ | $\left.\begin{gathered} n \\ \underset{\sim}{n} \end{gathered} \right\rvert\,$ | $\left\|\begin{array}{c} \text { ̇̀ } \end{array}\right\|$ | $\left\lvert\, \begin{gathered} \mathrm{i} \\ \hline \end{gathered}\right.$ | $\left\|\begin{array}{c} \infty \\ \text { di} \end{array}\right\|$ |  | $\left\|\begin{array}{c} \infty \\ \underset{\sim}{n} \end{array}\right\|$ | $\stackrel{n}{n}$ | $\overrightarrow{~ ̇ ~}$ | $\underset{\text { di}}{ }$ | $\underset{\sim}{n} \mid$ |  | $\underset{\sim}{n}$ | $\underset{N}{\mathrm{~N}}$ | $\left\lvert\, \begin{gathered} \infty \\ \text { di} \end{gathered}\right.$ | $\underset{\mathrm{i}}{ } \mid$ |  |  |  |  | $\underset{\sim}{\hat{j}}$ | $\stackrel{\ominus}{\mathrm{n}}$ |  |  |  |  |  |  | ה |  |
| 唇 | $\left\|\begin{array}{c} 0 \\ \vdots \\ י \end{array}\right\|$ | $\begin{array}{r} n \\ n \end{array}$ | $\left\|\begin{array}{l} 9 \\ \mathrm{I} \end{array}\right\|$ | $\hat{n}$ | $\underset{ \pm}{9}$ | $\begin{gathered} n \\ \end{gathered}$ | $\dot{m}$ |  | $\stackrel{\rightharpoonup}{i}$ |  | $\left\|\begin{array}{c} n \\ n \\ -1 \end{array}\right\|$ | $\underset{\sim}{\sim}$ | $\sim$ |  | $\stackrel{\sim}{\square}$ | $\pm$ | $\stackrel{3}{9}$ | $\stackrel{寸}{ \pm}$ |  | $\pm$ |  | $\stackrel{n}{ }$ |  | $\overline{\vec{n}}$ |  |  |  |  |  | － | $\stackrel{\bigcirc}{ \pm}$ |  |
| F | $\begin{array}{\|c\|} \hline n \\ \vdots \\ \hline \end{array}$ | $\stackrel{\circ}{6}$ | $\begin{array}{\|c\|} \hline n \\ \dot{\theta} \\ \hline \end{array}$ | ㅇ | $\begin{array}{\|c\|} \hline n \\ \vdots \\ \hline \end{array}$ | ¢ | \％ |  | N | $\stackrel{\circ}{\circ}$ | $\stackrel{\sim}{n}$ | $\stackrel{n}{0}$ | 알 |  | 6 | $\hat{0}$ | $\bigcirc$ | $\stackrel{n}{n}$ |  | － |  | $\gtrless$ | $\stackrel{\sim}{\sim}$ | 6 |  |  |  | 6 |  | ？ | ¢ |  |
| $\frac{80}{3}$ | $\begin{gathered} n \\ \end{gathered}$ | 앙 | $\begin{aligned} & n \\ & \vdots \\ & i \end{aligned}$ | 간 | $\begin{aligned} & n \\ & \vdots \\ & \vdots \end{aligned}$ | $\left\|\begin{array}{c} n \\ 0 \end{array}\right\|$ | \％ |  | N | $\stackrel{\circ}{\circ}$ | $\left\|\begin{array}{c} n \\ 0 \end{array}\right\|$ | 8 | $\stackrel{\sim}{\sim}$ |  | $n$ | $\hat{n}$ | $\bigcirc$ | $\begin{aligned} & n \\ & \vdots \\ & \hline \end{aligned}$ |  | － | 8 | N | $\underset{n}{n}$ | $\stackrel{n}{i}$ |  |  |  | ¢ |  | 8 | 3 |  |
| $\stackrel{\rightharpoonup}{0}$ | $\checkmark$ | z | $\lambda$ | $\rangle$ | $\rangle$ | $\rangle$ | $\rangle$ |  | $\lambda$ | $\lambda$ | $\rangle$ | Z | $\rangle$ |  | $\rangle$ | $\lambda$ | z | z | z | z | z | Z | z | $\lambda$ |  |  | z | z | ح | － | z |  |
| $\stackrel{\star}{\sim}$ | $\bigcirc$ | D | $D$ | $\bigcirc$ | $\bigcirc$ | D | $\bigcirc$ |  | $\bigcirc$ | $\sum_{\sum}^{\sum} \mid$ | $\bigcirc$ | $\sim$ | $\sum$ |  | 呺 | $\bigcirc$ | 堓 |  |  | － | 明 |  | $\bigcirc$ |  |  |  |  |  |  |  | $\bigcirc$ |  |
| $8$ | $\underset{4}{7}$ | $\underset{~}{7}$ | $\underset{~}{7}$ | $\underset{~}{7}$ | $\begin{array}{\|l\|} \underset{7}{7} \end{array}$ | $\underset{~}{7}$ | $\stackrel{\imath}{\imath}$ | 江 | $\left.\begin{array}{\|l\|} \ddagger \\ \vdots \end{array} \right\rvert\,$ | $\underset{~}{7}$ | $\underset{4}{\lambda}$ | $\underset{4}{7}$ | $\underset{4}{\Varangle}$ |  | $\underset{4}{~}$ | $\underset{4}{7}$ | $\underset{\sim}{7}$ | $\underset{4}{\lambda}$ |  | ¢ |  | $\underset{4}{㐅}$ | $\frac{\lambda}{4}$ | $\underset{4}{7}$ |  |  |  | ̇ |  | 近 | 江 | 交 |
|  | $\left\|\begin{array}{l} 4 \\ \vdots \\ \vdots \end{array}\right\|$ | $\left\|\begin{array}{l} \underset{a}{a} \\ \underset{\sim}{2} \end{array}\right\|$ | $\left.\begin{aligned} & \underset{\sim}{c} \\ & \underset{y}{4} \end{aligned} \right\rvert\,$ | $\left\|\begin{array}{c} \hat{\tilde{y}} \\ \underset{\sim}{n} \end{array}\right\|$ | $\|\underset{\sim}{\underset{\sim}{c}}\|$ | $\left.\begin{aligned} & \grave{j} \\ & \vdots \end{aligned} \right\rvert\,$ | $\left\|\begin{array}{c} \frac{2}{4} \\ \frac{2}{2} \end{array}\right\|$ | $\left\|\begin{array}{l} \underset{\sim}{2} \\ \frac{\pi}{0} \end{array}\right\|$ | $\left\|\begin{array}{l} 4 \\ \frac{1}{2} \end{array}\right\|$ | $\left\|\begin{array}{c} \underset{0}{0} \\ \hat{2} \end{array}\right\|$ | $\left\|\begin{array}{l} \underset{\sim}{2} \\ \frac{\pi}{0} \end{array}\right\|$ | $\stackrel{\rightharpoonup}{\infty}$ | $\begin{aligned} & \frac{2}{2} \\ & \frac{2}{n} \end{aligned}$ | $\left\|\begin{array}{c} \hat{0} \\ \hat{2} \\ \hline \end{array}\right\|$ | $\stackrel{\rightharpoonup}{4}$ | $\stackrel{\rightharpoonup}{\sim}$ | $\begin{aligned} & \underset{\sim}{4} \\ & \underset{\sim}{2} \end{aligned}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{\infty} \\ & \stackrel{\substack{c}}{ } \end{aligned}$ |  | 3 |  | $\stackrel{\substack{0 \\ \hline \\ \hline \\ \hline}}{ }$ |  | $\begin{aligned} & \stackrel{v}{\stackrel{2}{2}} \\ & \stackrel{\sim}{2} \end{aligned}$ |  | 3 |  | 3 |  | $\stackrel{\text { ¢ }}{4}$ | $\begin{aligned} & \text { 엉 } \\ & 2 \end{aligned}$ | $\frac{8}{4}$ |
| $\begin{array}{\|c\|c\|c\|c\|c\|c\|c\|c} \substack{50 \\ 0} \end{array}$ | $\left\|\begin{array}{c} \underset{\sim}{\wedge} \\ \underset{\sim y}{c} \end{array}\right\|$ | $\left\|\begin{array}{l} \stackrel{y}{\mathrm{~N}} \\ \stackrel{\sim}{c} \\ \hline \end{array}\right\|$ | $\left\|\begin{array}{c} \hat{\tilde{0}} \\ \hat{\imath} \mathrm{\alpha} \end{array}\right\|$ | $\left.\begin{array}{\|c} \frac{y}{c} \\ \underset{c}{c} \end{array} \right\rvert\,$ | $\left\|\begin{array}{l} 4 \\ \vdots \\ \text { 운 } \end{array}\right\|$ | $\left\|\begin{array}{c} \underset{\tilde{n}}{\hat{\sim}} \\ \hline \end{array}\right\|$ | $\left\|\begin{array}{c} 4 \\ \frac{1}{4} \\ \frac{y}{\infty} \end{array}\right\|$ | $\begin{array}{\|c\|c} \frac{y}{c} \\ \underset{y}{4} \end{array}$ | $\left\|\begin{array}{l} \vec{n} \\ \frac{2}{2} \end{array}\right\|$ | $\underset{\substack{e \\ \underset{4}{2} \\ \hline}}{ }$ | $\left\|\begin{array}{c} \frac{\mathfrak{\gamma}}{0} \\ \hdashline \end{array}\right\|$ | $\stackrel{\substack{\stackrel{y}{c} \\ 4 \\ 4}}{ }$ | $\begin{aligned} & \underset{\sim}{2} \\ & \underset{\sim}{2} \end{aligned}$ | $\left.\begin{aligned} & \underset{\sim}{c} \\ & \underset{y}{4} \end{aligned} \right\rvert\,$ | $\stackrel{a}{2}$ | $\underset{\sim}{\sim}$ | $\left.\begin{aligned} & \stackrel{\sim}{n} \\ & \underset{\sim}{n} \end{aligned} \right\rvert\,$ | $\frac{\stackrel{a}{2}}{4}$ |  | $\stackrel{y}{k}$ |  | $\begin{aligned} & \frac{\substack{0}}{3} \\ & \hline \end{aligned}$ | $\sum_{4}^{2}$ | $\begin{aligned} & 4 \\ & \frac{y}{9} \\ & \frac{y}{n} \end{aligned}$ |  |  |  | ＜ |  | ， | $\stackrel{\text { 줄 }}{4}$ | 2 |
| $\begin{gathered} \underline{x} \\ \vec{n} \\ \vec{n} \end{gathered}$ | ¢ | $\left\lvert\, \begin{gathered} \mathrm{g} \end{gathered}\right.$ | $$ | $\hat{\mathrm{t}} \mid$ | $\left\lvert\, \begin{gathered} \substack{~ \\ \downarrow} \end{gathered}\right.$ | $\mid \stackrel{g}{\mathrm{~d}}$ | 员 | $\begin{array}{\|c} 2 \\ \end{array}$ | $\hat{n}$ | $i n$ | $9$ | $\stackrel{5}{2}$ | $\underset{\substack{\mathrm{o} \\ \hline}}{ }$ | $\because$ | $\stackrel{?}{2}$ | $\stackrel{\sim}{2}$ | $\stackrel{\circ}{\circ}$ | $\stackrel{\square}{2}$ |  | 찬 |  | N | N | $\hat{\sim}$ |  | 찬 |  | $\infty$ |  | $\stackrel{2}{2}$ | $\stackrel{\text { ® }}{\sim}$ | $\stackrel{\sim}{2}$ |
| $\stackrel{x}{x}$ | $\left\|\begin{array}{c} 8 \\ \vdots \\ 4 \\ \sim \end{array}\right\|$ | $\left\lvert\, \begin{gathered} 8 \\ \substack{4 \\ ~} \end{gathered}\right.$ | $\left\|\begin{array}{c} 8 \\ \vdots \\ \underset{\sim}{2} \end{array}\right\|$ | $\left\lvert\, \begin{gathered} 8 \\ \substack{4 \\ ~} \end{gathered}\right.$ | $\left\|\begin{array}{c} o \\ \vdots \\ \underset{\sim}{2} \end{array}\right\|$ | $\left\lvert\, \begin{gathered} o \\ \underset{y}{c} \\ \hline \end{gathered}\right.$ | $\left\lvert\, \begin{gathered} 8 \\ \underset{y}{4} \\ \hline \end{gathered}\right.$ | $\left\|\begin{array}{c} 8 \\ d \\ d \end{array}\right\|$ | $\left\|\begin{array}{c} 8 \\ \vdots \\ \sim \end{array}\right\|$ | $\left\|\begin{array}{c} 8 \\ \underset{\sim}{2} \end{array}\right\|$ | $\left\|\begin{array}{c} 8 \\ \vdots \\ \sim \end{array}\right\|$ | $\stackrel{8}{8}$ | $\begin{aligned} & 8 \\ & \underset{\sim}{0} \end{aligned}$ | $\left\lvert\, \begin{gathered} 8 \\ \underset{\sim}{c} \\ \hline \end{gathered}\right.$ | $\begin{gathered} 0 \\ \vdots \\ \underset{y}{2} \end{gathered}$ | $\begin{gathered} 8 \\ \underset{\sim}{6} \end{gathered}$ | $\left\|\begin{array}{c} 8 \\ 0 \\ 4 \end{array}\right\|$ | $\begin{aligned} & 8 \\ & \vdots \\ & \underset{\sim}{2} \end{aligned}$ | 在 | $\begin{gathered} 0 \\ \underset{\sim}{4} \end{gathered}$ |  | $\begin{gathered} 0 \\ \underset{\sim}{4} \end{gathered}$ | $\begin{gathered} 8 \\ \underset{\sim}{6} \end{gathered}$ | $\begin{gathered} 8 \\ \vdots \\ \underset{y}{2} \end{gathered}$ |  | $\underset{\sim}{c}$ |  |  |  | $\stackrel{8}{4}$ | $\stackrel{8}{4}$ | $$ |
| $\begin{aligned} & 0 \\ & \text { تٍ } \\ & \text { ご } \end{aligned}$ |  | － | － | － | － | － | － | $\simeq$ | － | － | － | － | － | $\approx$ | － | － | $-$ | $\approx$ | － | － | － | $\simeq$ | － | $\approx$ |  | $\sim$ |  | － |  | $\simeq$ | － | $\approx$ |
| $\stackrel{\rightharpoonup}{\square}$ | ๆ | $\bigcirc$ | $\bigcirc$ | $\infty$ | $m$ | $a$ | $\infty$ | $\bigcirc$ | a | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | च | $\bigcirc$ | $\bigcirc$ | I | $\sim$ | $\sim$ |  | $\sim$ | m | in | － | $m$ |  | － |  | f |  | in | の |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\mathfrak{n}$ | 葡 | 4 |  |  |  | $\mathbb{4}$ | 宥 |  | に， |  | － |  |  |
| ธิ | $=$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\bigcirc$ | 2 | 2 | 2 | 2 | $\bigcirc$ | $\bigcirc$ | 9 | ন | $\infty$ | $\infty$ | $\infty$ | $\bigcirc$ |  | $a$ | $\sigma$ | $a$ |  | $\sigma$ |  | $\bigcirc$ |  | － | $\simeq$ | $\bigcirc$ |
| $\begin{aligned} & 5 \\ & \stackrel{y}{0} \\ & \stackrel{y}{0} \end{aligned}$ |  | － | － | － | － | － | － | － | － | ＋ | － | ＋ | － | ＋ | － | － | m | m |  | m |  | $m$ | m | $m$ |  | m |  | m |  | m | m |  |
| $\stackrel{\text { ® }}{\text { ¢ }}$ | $\left\|\begin{array}{c} \circ \\ \stackrel{\rightharpoonup}{\circ} \\ \hline \end{array}\right\|$ | $\begin{array}{\|c\|} \hline 0 \\ 0 \\ \text { N\| } \end{array}$ | $\left\lvert\, \begin{gathered} \circ \\ \stackrel{\rightharpoonup}{\circ} \\ \hline \end{gathered}\right.$ | $\begin{array}{\|c\|} \hline 0 \\ 0 \\ \text { N\| } \end{array}$ | $\left\lvert\, \begin{gathered} \circ \\ \stackrel{\rightharpoonup}{c} \\ \hline \end{gathered}\right.$ | $\left.\begin{array}{\|c\|} \hline 0 \\ \hline \\ \hline \end{array} \right\rvert\,$ | $\left\lvert\, \begin{gathered} \circ \\ \stackrel{\rightharpoonup}{\circ} \\ \hline \end{gathered}\right.$ | $\left\|\right\|$ | $\left\lvert\, \begin{gathered} \circ \\ \stackrel{\rightharpoonup}{\circ} \\ \hline \end{gathered}\right.$ | $\left.\begin{array}{\|c\|} \hline 0 \\ 0 \\ 1 \end{array} \right\rvert\,$ | $\left\|\begin{array}{c} \circ \\ \stackrel{\rightharpoonup}{0} \end{array}\right\|$ | $$ | $\stackrel{\circ}{\circ}$ | $\left\lvert\, \begin{gathered} \circ \\ \stackrel{\rightharpoonup}{\circ} \\ \hline \end{gathered}\right.$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \stackrel{\sim}{1} \end{aligned}$ | $\stackrel{\rightharpoonup}{0}$ | $\stackrel{\hat{N}}{\mathbf{N}}$ | － | 人 | $\stackrel{\circ}{i}$ | 颌 | $\stackrel{\rightharpoonup}{\mathbf{c}}$ | $\stackrel{\rightharpoonup}{8}$ |  | $\begin{array}{\|c\|} \hat{0} \\ \stackrel{\rightharpoonup}{2} \end{array}$ |  | $\begin{aligned} & \hat{\sim} \\ & \stackrel{\rightharpoonup}{0} \\ & \hline \end{aligned}$ |  | N | $\left\lvert\, \begin{gathered} \hat{o} \\ \mathbf{o} \\ \hline \end{gathered}\right.$ | － |




| Year | Month | Day | Location | Net | Capture | Prefix | Suffix | Leftleg | Rightleg | Age | Sex | Molt | Wing | Tail | Bill Length | Tarsus long | Tarsus short | $\begin{aligned} & \hline \mathrm{Wt} \\ & \text { bird } \end{aligned}$ | Pox status |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 3 | 14 | Airport Mitigation | 1 b | I | 2460 | 816 | GR/BL | PK/AL | AHY | FG | N | 64.5 | 65 | 13 | 22.1 | 23.2 | 10.5 | H |
| 2009 | 3 | 14 | Airport Mitigation | 3 | I | 2460 | 819 | GR/AL | GR/RD | AHY | FM | N | 65.5 | 65 | 11.8 | 23.4 | 24.8 | 11.5 | A |
| 2009 | 3 | 14 | Airport Mitigation | 3 | I | 2460 | 820 | BL/AL | BK/BL | HY | U | Y | 66.5 | 58 | 12.1 | 24.3 | 25.8 | 13.5 | H |
| 2009 | 3 | 15 | Airport Mitigation | 10 | I | 2460 | 822 | RD/WH | BK/AL | AHY | MG | N | 64 | 64 | 13.5 | 22.9 | 23.5 | 12.5 | H |
| 2009 | 3 | 15 | Airport Mitigation | 10 | I | 2460 | 823 | WH/GR | RD/AL | AHY | MG | N | 68.5 | 65 | 15.6 | 22.8 | 23.6 | 13 | H |
| 2009 | 3 | 15 | Airport Mitigation | 6 | R | 2460 | 727 | AL/BL | PK/BL | AHY |  | N | 68 | 63 | 14.7 | 21.1 | 22.3 | 12.5 | A? |
| 2009 | 3 | 15 | Airport Mitigation | 8b | R | 2460 | 964 | AL/RD | BK/WH | AHY |  | N | 69.5 | 65.5 | 13 | 23.8 | 25 | 12.5 | I? |
| 2009 | 3 | 15 | Airport Mitigation | 12b | I | 2460 | 825 | PK/WH | RD/AL | AHY | FG | N | 64 | 66.5 | 12.9 | 23.2 | 24.5 | 12 | H |
| 2009 | 3 | 15 | Airport Mitigation | 11 | I | 2460 | 826 | BL/AL | BL/WH | AHY | FG | N | 68.5 | 68 | 15.8 | 22.9 | 23.5 | 13.5 | H |
| 2009 | 3 | 15 | Airport Mitigation | 8b | R | 2460 | 734 | GR/BK | AL/PK | AHY |  | N | 67 | 64 | 15.5 | 22.8 | 23.5 | 12 | H |
| 2009 | 3 | 17 | Seaport TT | 4 | I | 2460 | 827 | AL/GR | BK/BL | AHY | FG | N | 68.5 | 65 | 13 | 24.6 | 25.9 | 12 | H |
| 2009 | 3 | 17 | Seaport TT | 1a | I | 2460 | 834 | GR/AL | RD/WH | HY | U | T | 67 | 66 | 14.1 | 22.4 | 23.2 | 13.5 | H |
| 2009 | 3 | 17 | Seaport TT | 4 | R | 2460 | 916 | AL/BL | WH/RD | AHY |  | N | 68 | 76 | 14.2 | 24.5 | 25.8 | 12 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 10 | I | 2460 | 853 | BL/AL | GR/PK | AHY | FG | N | 65.5 | 65.5 | 14.5 | 21.6 | 22.5 | 12 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 14 | I | 2460 | 854 | BL/PK | AL/BL | AHY | MG | N | 66 | 67 | 14.4 | 21.2 | 22.7 | 11 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 16 | I | 2460 | 857 | WH/RD | AL/GR | AHY | MG | N | 70 | 67 | 13.4 | 22.3 | 23.2 | 17 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 13 | I | 2460 | 859 | BK/GR | AL/BL | AHY | MG | N | 65.5 | 63 | 13 | 22.3 | 23.5 | 10.5 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 12 | I | 2460 | 860 | PK/BK | AL/GR | AHY | MG | N | 68.5 | 68 | 12.9 | 21.9 | 22.5 | 12 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 15 | I | 2460 | 861 | AL/PK | WH/BL | AHY | MG | N | 68 | 64 | 13.8 | 21.6 | 22.3 | 11.5 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 12 | I | 2460 | 865 | RD/AL | GR/BL | AHY | MG | N | 67.5 | 66 | 14.1 | 22.1 | 23 | 11.5 | H |
| 2009 | 3 | 18 | Seaport TT-Site 2 | 15 | I | 2460 | 870 | BK/BL | AL/PK | AHY | FG | N | 70.5 | 78.5 | 13.9 | 23.7 | 24.8 | 12 | H |
| 2009 | 3 | 20 | Santa Lourdes | 9a? | I | 2460 | 873 | AL/GR | BL/WH | AHY | FG | N | 68.5 | 65 | 14.1 | 23.4 | 23.9 | 12 | H |
| 2009 | 3 | 20 | Santa Lourdes | 16 | I | 2460 | 875 | AL/PK | BK/GR | AHY | MG | N | 67 | 65 | 14.2 | 22.6 | 23.2 | 9.2 | H? |
| 2009 | 3 | 22 | Seaport TT-Area 2 | 15a | I | 2460 | 876 | BL/AL | RD/BL | AHY | FM | N | 65.5 | 65 | 14.1 | 22.3 | 23.3 | 13 | H |
| 2009 | 3 | 22 | Seaport TT-Area 2 | 15b | I | 2460 | 877 | BK/AL | RD/PK | AHY | MG | N | 64.5 | 64 | 13.2 | 21.1 | 22.1 | 10.5 | H |
| 2009 | 3 | 22 | Seaport TT-Area 2 | 15c | I | 2460 | 878 | PK/WH | AL/RD | AHY | FG | N | 67.5 | 65.5 | 13.4 | 21.9 | 22.7 | 11.5 | H |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2010 | 3 | 11 | Seaport TT | T2 | I | 2460 | 883 | RD/BL | RD/AL | AHY | U | N | 71.0 | 69.0 | 14.1 | 23.0 | 22.1 | 12.5 | H |

Appendix B: Resights of Tinian Monarchs banded between 2006 and 2010 during post-delisting monitoring work for the Tinian Monarch. Color codes: $\mathrm{A}=$ Aluminum; $\mathrm{B}=\mathrm{Blue} ; \mathrm{R}=\mathrm{Red} ; \mathrm{G}=$ Green; P = Pink; K = Black; W = White. Age codes: AHY = After Hatch Year; HY = Hatch Year; $\mathrm{L}=$ Local (not fledged). Sex codes: $\mathrm{M}=$ Male; $\mathrm{F}=$ Female; $\mathrm{U}=$ Unknown. *Partial observation of band combination.

| Location | Band \# | Color Combo | Age | Sex | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Santa Lourdes | 00701 | ABBR | AHY | M | Banded | Observed | ---- | ---- | ---- |
| Santa Lourdes | 00702 | ARGB | AHY | M | Banded | Recaptured | Observed | Observed | ---- |
| Santa Lourdes | 00703 | AGRR | AHY | F | Banded | ---- | ---- | ---- | ---- |
| Santa Lourdes | 00705 | APGG | AHY | F | Banded | Recaptured | ---- | ---- | ---- |
| Santa Lourdes | 00706 | BAPP | AHY | F | Banded | Observed | Observed | Observed | Observed |
| Santa Lourdes | 00709 | GARB | AHY | M | Banded | ---- | ---- | ---- | ---- |
| Santa Lourdes | 00710 | RPAG | AHY | M | Banded | Recaptured | Observed | Observed | Observed |
| Santa Lourdes | 00711 | BGAP | AHY | F | Banded | Observed | Observed | Observed | Observed |
| Santa Lourdes | 00714 | GBAR | HY | U | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00716 | GRAG | AHY | M | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00717 | PGAB | AHY | F | Banded | Observed | Recaptured | Observed | Observed |
| Airport Mitigation | 00718 | BRAP | AHY | M | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00719 | GRAB | AHY | M | Banded | Observed | Recaptured | Observed | Observed |
| Airport Mitigation | 00723 | BAKR | AHY | M | Banded | Observed | Observed | Observed | Observed |
| Airport Mitigation | 00725 | AGBG | AHY | M | Banded | Recaptured | Observed | Observed | ---- |
| Airport Mitigation | 00726 | AKGR | HY | U | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00727 | ABPB | AHY | M | Banded | Recaptured | Observed | Recaptured | ---- |
| Airport Mitigation | 00732 | RAPG | AHY | F | Banded | Observed | Observed | Observed | Observed |
| Airport Mitigation | 00733 | PABK | AHY | M | Banded | Observed | Recaptured | Observed | Observed |
| Airport Mitigation | 00734 | GKAP | AHY | F | Banded | Observed | Observed | Recaptured | Observed |
| Airport Mitigation | 00735 | ARBR | AHY | M | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00736 | BGAK | AHY | F | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00737 | KAGP | AHY | M | Banded | Recaptured | Observed | Observed | ---- |
| Airport Mitigation | 00738 | GPPA | AHY | F | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00739 | GRBA | AHY | M | Banded | Observed | Recaptured | Observed | Observed |
| Santa Lourdes | 00742 | GPRA | AHY | M | Banded | Observed | Observed | Observed | Observed |
| Santa Lourdes | 00746 | RGAR | AHY | M | Banded | Observed | Recaptured | Observed | Observed |
| Santa Lourdes | 00747 | APBG | AHY | M | Banded | Observed | Observed | ---- | ---- |
| Santa Lourdes | 00748 | GABR | AHY | F | Banded | ---- | ---- | ---- | ---- |
| Santa Lourdes | 00749 | BGBA | AHY | F | Banded | ---- | ---- | ---- | ---- |
| Santa Lourdes | 00750 | KAPP | AHY | F | Banded | Observed | Observed | Observed | Observed |
| Airport Mitigation | 00753 | PBPA | AHY | M | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00755 | ARRG | AHY | M | Banded | Observed | Observed* | ---- | ---- |
| Airport Mitigation | 00760 | AGGR | AHY | M | Banded | Recaptured | Observed | Observed | Observed |
| Airport Mitigation | 00761 | AKRB | AHY | F | Banded | ---- | ---- | ---- | ---- |
| Airport Mitigation | 00762 | GAKP | AHY | M | Banded | Observed | Observed | Observed | Observed |
| Airport Mitigation | 00763 | RRAB | AHY | F | Banded | Observed | Observed | ---- | ---- |


| Location | Band \# | Color <br> Combo | Age | Sex | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Santa Lourdes | 00765 | ABGB | AHY | F | Banded | ---- | ---- | ---- | ---- |
| Santa Lourdes | 00769 | BBGA | AHY | F |  | Banded | ---- | ---- | ---- |
| Santa Lourdes | 00770 | KRBA | AHY | F |  | Banded | ---- | ---- | ---- |
| Airport Mitigation | 00772 | APWW | AHY | M |  | Banded | Recaptured | Observed | Observed |
| Airport Mitigation | 00774 | AWRR | AHY | F |  | Banded | Recaptured | ---- | ---- |
| Airport Mitigation | 00776 | APWR | AHY | F |  | Banded | Observed | Observed | Observed |
| Santa Lourdes | 00782 | ABWG | HY | U |  | Banded | ---- | ---- | ---- |
| Airport Mitigation | 00785 | GWRA | AHY | M |  | Banded | Observed | ---- | ---- |
| Santa Lourdes | 00789 | AGPG | HY | U |  | Banded | ---- | ---- | ---- |
| Airport Mitigation | 00792 | RRBA | AHY | F |  | Banded | Observed | Observed | Observed |
| Santa Lourdes | 00797 | ARGW | AHY | M |  | Banded | Observed | Observed | Observed |
| Santa Lourdes | 00799 | ABRW | AHY | F |  | Banded | -- | ---- | ---- |
| Santa Lourdes | 00803 | AGGB | AHY | F |  |  |  | Banded | Observed |
| Santa Lourdes | 00804 | ARWB | AHY | F |  |  |  | Banded | Observed |
| Santa Lourdes | 00805 | RAGK | AHY | M |  |  |  | Banded | ---- |
| Airport Mitigation | 00809 | APKW | AHY | F |  |  |  | Banded | Observed |
| Airport Mitigation | 00811 | PKRA | AHY | F |  |  |  | Banded | Observed |
| Airport Mitigation | 00812 | KGPA | AHY | F |  |  |  | Banded | Observed |
| Airport Mitigation | 00813 | BAWG | AHY | F |  |  |  | Banded | ---- |
| Airport Mitigation | 00814 | RGAP | AHY | M |  |  |  | Banded | Observed |
| Airport Mitigation | 00816 | GBPA | AHY | F |  |  |  | Banded | ---- |
| Airport Mitigation | 00819 | GAGR | AHY | F |  |  |  | Banded | ---- |
| Airport Mitigation | 00820 | BAKB | HY | U |  |  |  | Banded | ---- |
| Airport Mitigation | 00822 | RWKA | AHY | M |  |  |  | Banded | Observed |
| Airport Mitigation | 00823 | WGRA | AHY | M |  |  |  | Banded | ---- |
| Airport Mitigation | 00825 | PWRA | AHY | F |  |  |  | Banded | Observed |
| Airport Mitigation | 00826 | BABW | AHY | F |  |  |  | Banded | ---- |
| Seaport TT | 00827 | AGKB | AHY | F |  |  |  | Banded | Observed |
| Seaport TT | 00834 | GARW | HY | U |  |  |  | Banded | ---- |
| Seaport TT | 00853 | BAGP | AHY | F |  |  |  | Banded | ---- |
| Seaport TT | 00854 | BPAB | AHY | M |  |  |  | Banded | Observed |
| Seaport TT | 00857 | WRAG | AHY | M |  |  |  | Banded | Observed |
| Seaport TT | 00859 | KGAB | AHY | M |  |  |  | Banded | Observed |
| Seaport TT | 00860 | PKAG | AHY | M |  |  |  | Banded | ---- |
| Seaport TT | 00861 | APWB | AHY | M |  |  |  | Banded | ---- |
| Seaport TT | 00865 | RAGB | AHY | M |  |  |  | Banded | Observed |
| Seaport TT | 00870 | KBAP | AHY | F |  |  |  | Banded | Observed |
| Santa Lourdes | 00873 | AGBW | AHY | F |  |  |  | Banded | Observed |
| Santa Lourdes | 00875 | APKG | AHY | M |  |  |  | Banded | Observed |
| Seaport TT | 00876 | BARB | AHY | F |  |  |  | Banded | Observed |
| Seaport TT | 00877 | KARP | AHY | M |  |  |  | Banded | Observed |
| Seaport TT | 00878 | PWAR | AHY | F |  |  |  | Banded | Observed |
| Seaport TT | 00883 | RBRA | AHY | U |  |  |  |  | Banded |
| Santa Lourdes | 00902 | KWPA | AHY | M |  | Banded | ---- | ---- | ---- |


| Location | Band \# | Color <br> Combo | Age | Sex | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Santa Lourdes | 00903 | WGAG | AHY | F |  | Banded | Observed | ---- | ---- |
| Santa Lourdes | 00904 | ARKG | AHY | F |  | Banded | Observed | Observed | Observed |
| Santa Lourdes | 00905 | RBAB | AHY | M |  | Banded | Observed | Observed | Observed |
| Santa Lourdes | 00906 | PKAP | L | U |  | Banded | ---- | ---- | ---- |
| Santa Lourdes | 00907 | KBAW | L | U |  | Banded | ---- | ---- | ---- |
| Airport Mitigation | 00908 | RARB | AHY | M |  | Banded | Observed | --- | ---- |
| Airport Mitigation | 00909 | RKPA | AHY | F |  | Banded | Observed | Observed | Observed |
| Airport Mitigation | 00911 | BRGA | AHY | M |  | Banded | Observed | ---- | ---- |
| Airport Mitigation | 00912 | GAGB | AHY | F |  | Banded | Observed | ---- | ---- |
| Airport Mitigation | 00914 | PKAK | L | U |  | Banded | ---- | ---- | ---- |
| Airport Mitigation | 00915 | KWAW | L | U |  | Banded | ---- | ---- | ---- |
| Seaport TT | 00916 | ABWR | AHY | M |  | Banded | Observed | Recaptured | ---- |
| Seaport TT | 00917 | AWBG | AHY | M |  | Banded | ---- | ---- | Observed |
| Seaport TT | 00918 | GBAK | AHY | F |  | Banded | Observed | ---- | ---- |
| Seaport TT | 00919 | WBAP | AHY | M |  | Banded | Observed | Observed | ---- |
| Seaport TT | 00923 | ARRB | AHY | F |  | Banded | ---- | ---- | ---- |
| Santa Lourdes | 00924 | ABGK | AHY | F |  |  | Banded | -- | ---- |
| Santa Lourdes | 00927 | APPW | U | F |  |  | Banded | Observed |  |
| Santa Lourdes | 00940 | APGR | AHY | F |  |  | Banded | Observed | Observed |
| Santa Lourdes | 00942 | KPRA | AHY | F |  |  | Banded | ---- | ---- |
| Santa Lourdes | 00943 | ABRP | AHY | F |  |  | Banded | Observed | Observed |
| Santa Lourdes | 00945 | RWPA | HY | U |  |  | Banded | Observed |  |
| Airport Mitigation | 00950 | PWBA | HY | U |  |  | Banded | ---- | ---- |
| Airport Mitigation | 00951 | AGPW | AHY | M |  |  | Banded | Observed | Observed |
| Airport Mitigation | 00952 | AWPB | AHY | F |  |  | Banded | ---- | Observed |
| Airport Mitigation | 00953 | AWKW | AHY | M |  |  | Banded | ---- | ---- |
| Airport Mitigation | 00954 | GARP | AHY | M |  |  | Banded | -- | ---- |
| Airport Mitigation | 00964 | ARKW | AHY | M |  |  | Banded | Recaptured | Observed |
| Seaport TT | 00967 | APBP | AHY | M |  |  | Banded | Observed | Observed |
| Seaport TT | 00972 | AWRK | HY | U |  |  | Banded | ---- | ---- |
| Seaport TT | 00974 | BABP | AHY | F |  |  | Banded | Observed | Observed |
| Seaport TT | 00975 | RABW | HY | U |  |  | Banded | -- | ---- |
| Seaport TT | 00976 | WAKB | AHY | F |  |  | Banded | Observed | Observed |
| Seaport TT | 00977 | AWWP | AHY | M |  |  | Banded | Observed | Observed |
| Seaport TT | 00978 | AKKP | AHY | F |  |  | Banded | Observed | ---- |
| Seaport TT | 00979 | AWPW | AHY | F |  |  | Banded | Observed | Observed |
| Seaport TT | 00980 | BAKW | AHY | M |  |  | Banded | Observed | Observed |
| Seaport TT | 00981 | WPAK | AHY | F |  |  | Banded | ---- | ---- |
| Seaport TT | 00997 | WBAW | AHY | M |  |  | Banded | Observed | Observed |

Appendix C: Tinian Monarch territory mapping figures from the 2007-2011 annual monitoring reports.


Map 1. Estimated locations of Tinian monarch territories in the Santa Lourdes study site in March 2007.


Map 2. Estimated locations of Tinian monarch territories in the Airport Mitigation study site in March 2007.


Map 3. Estimated locations of Tinian monarch territories in the Santa Lourdes study site in March 2008.


Map 4. Estimated locations of Tinian monarch territories in the Airport Mitigation Area study site in March 2008.


Map 5. Estimated locations of Tinian monarch territories in the Seaport Tangantangan study site in March 2008.


Map 6. Estimated locations of Tinian monarch territories in the Santa Lourdes study site in March 2009.


Map 7. Estimated locations of Tinian monarch territories in the Airport Mitigation study site in March 2009.


Map 8. Estimated locations of Tinian monarch territories in the Seaport Tangantangan study site in March 2009.


Map 9. Estimated locations of Tinian monarch territories in the Santa Lourdes study site in March 2010.


Map 10. Estimated locations of Tinian monarch territories in the Airport Mitigation study site in March 2010.


Map 11. Estimated locations of Tinian monarch territories in the Seaport Tangantangan study site in March 2010.


Map 12. Estimated locations of Tinian monarch territories in the Santa Lourdes study site in March 2011.


Map 13. Estimated locations of Tinian monarch territories in the Seaport Tangantangan study site in March 2011.


[^0]:    ${ }^{1}$ U.S. Fish and Wildlife Service, Pacific Islands Fish and Wildlife Office, 300 Ala Moana Boulevard, Room 3-122, Honolulu, Hawaii 96850
    ${ }^{2}$ Division of Fish and Wildlife, Department of Lands and Natural Resources, P.O. Box 10007, Saipan, MP 96950
    ${ }^{3}$ Pacific Rim Conservation, Honolulu, Hawaii
    ${ }^{4}$ Hawaii Pacific University, Honolulu, Hawaii

